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Atmospheric deposition of microplastics in Shiraz, Iran

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Abstract

 The atmosphere plays a critical role in the regional and global transportation and redistribution of microplastics (MPs). However, the significance of rainfall and its means of scavenging MPs are not well understood. In this study, MP deposition was determined during successive dry and rainy events over eight consecutive days in the Shiraz region of Iran. Flux magnitudes and temporal distributions at six sites within and outside the city (including a remote, non- urbanised location) were similar and revealed a progressive increase in MP abundance and 55 deposition during dry periods (up to about 50 MP m^{-2} h⁻¹) and subsequent relatively rapid wet 56 deposition (washout) by incipient rainfall (peaking at about 130 MP m^{-2} h⁻¹). Wet deposition of MPs progressively decreased throughout the rainfall event, but with evidence of secondary peaks, before atmospheric accumulation and dry deposition increased during the next dry event and the cycle continued. Dry and wet deposition were dominated by fibres (that included polyester-polyethylene terephthalate, polystyrene, polyvinyl chloride and polyethylene) but the evolution of deposition did not appear to be associated with changes in MP size. These observations indicate that, as with other airborne pollutants, initial rainfall is an efficient scavenger of atmospheric MPs. Along with variations in methodology, this effect may contribute to the wide variation in MP fluxes reported in the literature.

- **Keywords:** fibres; accumulation; precipitation; airborne; scavenging
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1. Introduction

 It is becoming increasingly evident that the atmosphere is a key temporary reservoir for microplastics (MPs; < 5 mm in size) (Dris et al., 2016; Wright et al., 2020; Zhang et al., 2020;). As a consequence, MPs, and in particular those of a fibrous nature, may be transported long distances from their direct sources and have been detected in regions remote from centres of urbanisation and industrialisation (Allen et al., 2019; Brahney et al., 2020; Ding et al., 2021; Villanova-Solano et al., 2023).

 The significance of airborne MPs is often quantified in terms of dry, wet or bulk deposition; that is, near-ground level deposition rate over a specific area and usually expressed as number per $\frac{m^2}{2}$ per day. On this basis, independent results have revealed variations spanning many orders of magnitude (Purwiyanto et al., 2022; Li et al., 2023). Part of this variation may be attributed to differences in sampling, processing and analytical detection limits and distances from known sources, but environmental factors are also important. In particular, rainfall is believed to wash out MPs from the atmosphere by wet deposition, and it has been suggested that the efficacy of this effect may be related to rainfall intensity and duration and rain droplet size (Szewc et al., 2021; Yuan et al., 2023). However, the precise means and timeframes involved in MP scavenging by rainfall are not well understood, largely because most studies have considered aggregated deposition or net accumulation over extended periods or entire events (Abbasi and Turner, 2021; Huang et al., 2021; Li et al., 2023).

 In a short-term study of a monsoon event, Abbasi (2021) showed that the wet depositional flux of MPs decreased rapidly in the first 30 minutes, providing evidence for the rapid washing out or airborne MPs during seasonal, heavy rainfall. In the present study, the more general role of rainfall in scavenging MPs from the atmosphere was investigated over a longer timeframe that encompasses successive dry and wet events with the hypothesis that wet deposition is not consistent throughout an event. Specifically, dry and wet (or bulk) MP deposition were determined over 24 periods of three to twelve hours each and at six different locations in the region of Shiraz, Iran. MPs were also classified by size and shape in order to evaluate whether these characteristics are impacted by environmental conditions (rainfall, wind speed, humidity)

 and a selection was analysed for their polymer composition in order to identify the principal plastic types.

2. Material and Methods

2.1. Sample location

 Shiraz, southwest Iran, is the fifth most populous city in Iran (about 1.6 million in 2016) and lies on a green plain at about 1600 m above sea level. The city supports an oil refinery and various industries in the electronics, manufacturing and agriculture sectors. The climate is moderate 105 semi-arid with an annual average rainfall, humidity and temperature of 335 mm, 64% and 18 °C, respectively, and the prevailing wind direction is from the northwest. Samples were collected from the urban metropolis of Shiraz (S1 to S5) and from a station remote from any significant urbanisation (C) located about 40 km to the north and upwind of the city (Figure 1). Regarding the Shiraz sample locations, S1 is a low traffic area on the outskirts of the city and close to agricultural land to the west, S2 and S3 are residential areas of low traffic and a high density of green space, S4 is the northern entrance to the city centre and is adjacent to a major highway, and S5 is a residential area with a moderate amount of traffic.

114 Figure 1: Location of the sampling sites in the Shiraz region. Note the relative remoteness of **station C from centres of population (in grey).**

2.2. Sampling

 After carefully checking the rainfall forecasts, eight consecutive days (192 h) were selected for sampling from 28 December 2021 to 5 Jan 2022. This timescale comprised three dry events (totalling 117 h) and two distinct rainfall regimes (totalling 75 h). Dry events were sampled across 11 periods (labelled D) of twelve hours (plus any remaining time before incipient rain), and rain

 events were sampled across 13 periods (labelled R) of three to eight hours depending on the duration of precipitation.

 Samples were collected at each location in a stainless steel dish (whose depth was 30 cm and aperture diameter ranged from 25 to 43 cm, depending on any constraints imposed by the precise sampling location and position). Each dish was positioned, via a metallic frame, about 2 m above the flat roof of a three-storey building that was exposed in all directions, and at a total elevation of about 10 m above ground level. During dry events, dishes were partly filled with 1-L 129 of 2 µm-filtered, distilled water to aid retention and minimise blowby of dry deposited material. 130 At each site, material collected over the sampling period, plus filtered $(< 2 \mu m)$, distilled water- rinsings of the dish, were transferred to a 2-L glass bottle with the aid of a glass funnel. During rain events, pre-cleaned (with distilled water) dry dishes were continuously used to collect wet deposited material over each period (although, strictly, bulk deposition was captured). The contents were carefully poured into a graduated glass measuring cylinder to determine rainfall volume. With a Eutech Instruments PCD650 probe, electrical conductivity (as specific conductance and a measure of rainwater chemistry) was then determined before the contents were transferred, with the dish rinsings, to a glass bottle as above.

2.3. Sample processing and microplastic counting and characterisation

 The contents of each glass sample bottle were vacuum filtered through a 2 μm pore size S&S filter paper (Blue Band, grade 589/3). Filters were transferred to individual, 60 mL foil-wrapped 142 glass beakers and digested in 35 mL of 30% H_2O_2 (Arman Sina, Tehran) until bubble formation

143 ceased. The contents of the digestions were refiltered through clean 2 μ m filters that were 144 subsequently air-dried at 25 \degree C for 48 h in a metal cabinet housed in a clean room before being stored in glass Petri dishes.

 MPs on each filter were visually identified, counted and characterised (by size, shape and colour) under a stereo digital microscope (Sairan DSM3000) at up to 200 X magnification with the aid of 148 a 250 µm-diameter stainless steel probe and ImageJ software. Identification of MPs was based on shape, colour, size, thickness, shininess, hardness and surface structure according to protocols outlined elsewhere (Abbasi et al., 2019). Classification was based on shape (fibre, film, fragment or spherule) and length of the longest axis, *L* (*L* < 100 µm, 100 ≤ *L* < 250 µm, 250 ≤ *L* < 500 µm, 500 ≤ *L* < 1000 µm, *L* > 1000 µm; and with a size detection limit estimated to be between 30 and 153 50 μ m depending on shape). The polymeric composition of 34 MPs (or about 4% of the total) of a range of shapes, sizes and colours and collected from different locations and meteorological conditions was determined using a micro-Raman spectrometer (µ-Raman-532-Ci, Avantes, 156 Apeldoorn, Netherland) with a laser of 785 nm and Raman shift of 400-1800 $cm⁻¹$.

2.4. Laboratory cleanliness and quality assurance

 Laboratory equipment was washed with phosphate-free soap, double-rinsed with distilled water 159 and soaked in 10% HNO₃ for 24 h before being rinsed twice with filtered, distilled water, dried at room temperature in a customised clean room and protected by Al foil. Laboratory work surfaces were cleaned with ethanol, laboratory clothing was cotton-based and all reagents and solutions were filtered through 2 μm before being used. Under these conditions, processing of distilled water contained in glass bottles as above revealed no MP contamination. The number of MPs on

 ten random (sample) filters were recounted under the microscope and returned the same values as the original counts.

3. Results

 Table 1 shows the duration of each sampling event, along with an indication of regional wind speed, wind direction and relative humidity (obtained from ventusky.com) for each period. 170 Average wind speeds ranged from $<$ 5 to 10 - 15 km h⁻¹ and usually had a southerly and/or westerly component. Relative humidity exceeded 80% during the rain events but was more variable during dry events (between 10% and 90%). Also shown for each rainy period is rainfall (in mm), derived from the volume of precipitation collected and the area of the metal collecting dish, along with the measured specific electrical conductance of rainwater. For a given period, rainfall exhibited some variation among the different locations, and in particular the highest values were frequently reported for S4 to the north of the city. Overall, average rainfall intensity (i.e., mean periodic rainfall for all locations normalised to sampling duration) ranged from about 0.25 mm h⁻¹ for R6 to > 2.2 mm h⁻¹ for R18, and specific electrical conductance, as a measure of 179 rainwater chemistry, ranged from about 20 to 270 μ s cm⁻¹, with the highest values at each location usually encountered towards the beginning of both rain events.

 Table 2 presents the number of MPs retrieved from the filters at each location and for each 182 sampling period, along with the number normalised per $m²$ of sampling area and per h of collection time (i.e., deposition rate). In total, 132 samples were collected and analysed (twelve were unsuccessfully retrieved or lost), and in all but eleven of these samples MPs were identified.

 The number of MPs in each sample was > 20 in eleven cases, and the maximum number of MPs 186 was 47. When normalised, deposition rates (where detected) ranged from < 1 m⁻² h⁻¹ (< 24 m⁻² d⁻ -1) in three samples taken during dry conditions to > 100 m⁻² h⁻¹ (> 2400 m⁻² d⁻¹) in three samples collected during rain events.

190 Table 1: Duration of the dry (D) and rainy (R) sampling events and periods, along with indicative, average (by rank) regional wind speeds and directions and

- 191 relative humidities. Periodic rainfall, derived from the volume of precipitation and area of sample capture, and specific electrical conductance (EC) of rainwater,
- 192 are shown for each location. Note that NC denotes no sample collection and LS denotes a lost sample.

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195 Table 2: The number of MPs and the number of MPs per m² per h for each sampling event and sampling period and for each location. Note that NC denotes no

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 Figure 2: MP deposition as a function of time at the six locations (C and S1 to S5). Dry events are shown in red and rain events are shown in blue. Note the discontinuities where samples were lost or not collected (see caption to Figure 2).

- 249 Figure 3: Percentage distribution of MPs by size (*L*, in μ m) for five different dry (D) and rain (R)
- events (see Table 1).

 Figure 4: Distribution of MPs by polymer type (*n* = 34). PET = polyethylene terephthalate, PTFE = polytetrafluoroethylene, and numbers in parentheses in the legend denote non-fibrous samples (fragments).

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4. Discussion

 Based on our sampling design and identification protocols, the broad, qualitative findings of the present study are in agreement with those from other studies examining the presence or 267 deposition of atmospheric MPs in urban and more remote areas, and either in dry conditions or during rainfall or both. Here, fibres of varying lengths are the most abundant shape of MP, presumably because of their favourable aerodynamic properties, and PET-polyester is the most common (or one of most common) petroleum-based polymer/s (Wright et al., 2020; Abbasi and Turner, 2021; Ding et al., 2021; Huang et al., 2021; Purwiyanto et al., 2021; Jia et al., 2022; Liu et al., 2022; Perera et al., 2022; Chang et al., 2023; Li et al., 2023). The dominance of polyester-273 based fibres suggest that clothing and textiles are the most important primary source of airborne MPs, with secondary distal sources likely to include resuspension from soils and agricultural land applied with contaminated sewage sludge (Brahney et al., 2020; Rezaei et al., 2022).

Atmospheric deposition rates of MPs reported in the literature vary widely. For example,

278 Purwiyanto et al. (2022) determined fluxes between 3 and 40 MPs m^{-2} d⁻¹ in coastal Jakarta,

279 while Jia et al. (2022) report an average of 3261 MPs m⁻² d⁻¹ in Shanghai during the rainy season 280 and Li et al. (2023) report a maximum of 75,000 MPs m⁻² d⁻¹ in a rural region of northern China. Geographical and temporal variations have been attributed to seasonal and climatic conditions (Huang et al., 2021; Hernández-Fernández et al., 2022) and proximity to urban areas (Shruti et al., 2022; Sun et al., 2022), as well as differences in sample collection (including height above ground level) and processing and means of plastic identification (Knobloch et al., 2021; Szewc et al., 2021). For example, we note that Li et al. (2023) sampled at only 100 cm above ground level and report rayon as the dominant polymer type, indicating that semi-synthetic fibres were

considered alongside petroleum-based polymers.

One of the key factors determining the nature and magnitude of atmospheric MP depositional

flux is the mode of deposition (dry deposition versus wet deposition). Dry deposition is

determined by deposition velocity (particle size and shape) and is also specific to

meteorological conditions and landcover type (Klein et al., 2019; Szewc et al., 2021), while

 rainfall acts to scavenge MPs from the atmosphere (Zhang et al., 2020). Many studies have found that, for a given sampling time, wet deposition is greater than dry deposition, or at least bulk deposition is greater during wet periods than dry periods (Dris et al., 2016; Huang et al., 2021; Liu et al., 2022; Li et al., 2023; Yuan et al., 2023). However, these studies have only considered daily or cumulative rainfall and have acknowledged that other factors, like rainfall height, intensity and frequency and droplet size, could be important. The potential effects of rainfall intensity and duration on MP deposition were demonstrated by Abbasi (2021) during the onset of a monsoon event in Shiraz. Here, MP deposition rate rapidly declined during the first 30 minutes of heavy precipitation.

 The present study has determined MP deposition with a relatively high temporal resolution 302 (several hours, and with MPs reported per $m⁻²$ per h) over consecutive dry and rain events and, therefore, affords a greater insight into the role that rainfall plays in MP transport. Thus, across six locations within the vicinity of Shiraz that includes a more remote location, and despite different land uses and local sources of MPs, distributions and magnitudes of deposition across the entire sampling timeframe are similar.

 Specifically, an initial dry weather event, with southerly winds and low humidity, allows MPs from local and distal sources to progressively build up in the atmosphere and deposit through gravitational settlement. Incipient rainfall acts to rain out and wash out these MPs (through in- cloud and below-cloud scavenging, respectively; Audoux et al., 2023), resulting in a spike in wet deposition with little or no lag. Subsequent precipitation encounters "cleaner" air and depositional fluxes decline to a low, residual value. A secondary peak in deposition could be related to a spell of relatively high rainfall (and, presumably, relatively large raindrops that increase collision frequencies; Guo et al., 2016) coupled with winds with an easterly component (Table 1) that introduce MPs from a different region. Specific electrical conductance follows these broad trends, reflecting the propensity of rainfall to concurrently washout airborne salts from the atmosphere.

 The ensuing dry event is accompanied by an immediate period of low deposition (initially, no MPs were detected at C and S1), and although there is often a subsequent build-up of MPs, the

 increase is not as marked as that observed during the initial dry event. Presumably, this reflects the effects of preceding rainfall in cleansing atmospheric MPs and the wetting of regional soils that inhibits MP resuspension (Abbasi and Turner, 2021). An additional contributing factor could be that rainfall here began in the early hours of the next day when local MP-generating activities were minimal. The next rainfall event is accompanied by a smaller peak that reflects the washing out of lower numbers of locally resuspended and regionally advected MPs, with the final dry event beginning with very little MP deposition and exhibiting evidence that the preceding cycle is repeated.

 The intra- and inter-event patterns that we observed for MPs, and in particular those during incipient rainfall, have been documented for other aerosols more generally (Battarbee et al., 1997; Castro et al., 2010; Fujino and Miyamoto, 2022). Here, washout is usually reported to be greater for larger particles, and although this effect was not evident from our results, it must be borne in mind that the more general literature deals with much smaller and higher density particulate matter. Consistent with other independent studies, rainfall acts to remove MPs from the atmosphere, at least temporarily.

 In summary, average MP depositional fluxes were similar among multiple locations with different characteristics within the Shiraz region, and significant (at least order of magnitude) intra-location variations observed during different meteorological conditions. These findings support previous assertions that MPs are ubiquitous contaminants, but also suggest that variations reported in the literature may result, to a significant extent, from environmental conditions and not, necessarily, proximity to source regions. Specifically, it would appear that incipient rainfall is a critical driver for the scavenging of MPs from the atmosphere through below-cloud washout. It is surmised that rainfall also acts to inhibit MP resuspension from certain surfaces (e.g., soils) through ground dampening.

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