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# Atmospheric deposition of microplastics in Shiraz, Iran

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35	Highlights
36	Atmospheric microplastics (MPs) accumulate during dry weather over Shiraz
37	Subsequently, MPs rapidly wash out with incipient rain
38	Deposition decreases and is lowest when rainfall ceases and the cycle continues
39	MPs dominated by fibres but no clear relationship between deposition mode and fibre size
40	Variability of MP deposition in the literature may reflect environmental conditions
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#### Abstract

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The atmosphere plays a critical role in the regional and global transportation and redistribution of microplastics (MPs). However, the significance of rainfall and its means of scavenging MPs are not well understood. In this study, MP deposition was determined during successive dry and rainy events over eight consecutive days in the Shiraz region of Iran. Flux magnitudes and temporal distributions at six sites within and outside the city (including a remote, nonurbanised location) were similar and revealed a progressive increase in MP abundance and deposition during dry periods (up to about 50 MP m<sup>-2</sup> h<sup>-1</sup>) and subsequent relatively rapid wet deposition (washout) by incipient rainfall (peaking at about 130 MP m<sup>-2</sup> h<sup>-1</sup>). Wet deposition of MPs progressively decreased throughout the rainfall event, but with evidence of secondary peaks, before atmospheric accumulation and dry deposition increased during the next dry event and the cycle continued. Dry and wet deposition were dominated by fibres (that included polyester-polyethylene terephthalate, polystyrene, polyvinyl chloride and polyethylene) but the evolution of deposition did not appear to be associated with changes in MP size. These observations indicate that, as with other airborne pollutants, initial rainfall is an efficient scavenger of atmospheric MPs. Along with variations in methodology, this effect may contribute to the wide variation in MP fluxes reported in the literature.

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**Keywords:** fibres; accumulation; precipitation; airborne; scavenging

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#### 1. Introduction

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71 It is becoming increasingly evident that the atmosphere is a key temporary reservoir for 72 microplastics (MPs; < 5 mm in size) (Dris et al., 2016; Wright et al., 2020; Zhang et al., 2020;). As 73 a consequence, MPs, and in particular those of a fibrous nature, may be transported long 74 distances from their direct sources and have been detected in regions remote from centres of 75 urbanisation and industrialisation (Allen et al., 2019; Brahney et al., 2020; Ding et al., 2021; 76 Villanova-Solano et al., 2023). 77 The significance of airborne MPs is often quantified in terms of dry, wet or bulk deposition; that is, near-ground level deposition rate over a specific area and usually expressed as number per 78 79 m<sup>2</sup> per day. On this basis, independent results have revealed variations spanning many orders 80 of magnitude (Purwiyanto et al., 2022; Li et al., 2023). Part of this variation may be attributed to 81 differences in sampling, processing and analytical detection limits and distances from known 82 sources, but environmental factors are also important. In particular, rainfall is believed to wash 83 out MPs from the atmosphere by wet deposition, and it has been suggested that the efficacy of 84 this effect may be related to rainfall intensity and duration and rain droplet size (Szewc et al., 85 2021; Yuan et al., 2023). However, the precise means and timeframes involved in MP 86 scavenging by rainfall are not well understood, largely because most studies have considered 87 aggregated deposition or net accumulation over extended periods or entire events (Abbasi and 88 Turner, 2021; Huang et al., 2021; Li et al., 2023). 89 In a short-term study of a monsoon event, Abbasi (2021) showed that the wet depositional flux 90 of MPs decreased rapidly in the first 30 minutes, providing evidence for the rapid washing out 91 or airborne MPs during seasonal, heavy rainfall. In the present study, the more general role of 92 rainfall in scavenging MPs from the atmosphere was investigated over a longer timeframe that 93 encompasses successive dry and wet events with the hypothesis that wet deposition is not consistent throughout an event. Specifically, dry and wet (or bulk) MP deposition were 94 95 determined over 24 periods of three to twelve hours each and at six different locations in the region of Shiraz, Iran. MPs were also classified by size and shape in order to evaluate whether 96 97 these characteristics are impacted by environmental conditions (rainfall, wind speed, humidity)

and a selection was analysed for their polymer composition in order to identify the principal plastic types.

# 2. Material and Methods

# 2.1. Sample location

Shiraz, southwest Iran, is the fifth most populous city in Iran (about 1.6 million in 2016) and lies on a green plain at about 1600 m above sea level. The city supports an oil refinery and various industries in the electronics, manufacturing and agriculture sectors. The climate is moderate semi-arid with an annual average rainfall, humidity and temperature of 335 mm, 64% and 18 °C, respectively, and the prevailing wind direction is from the northwest. Samples were collected from the urban metropolis of Shiraz (S1 to S5) and from a station remote from any significant urbanisation (C) located about 40 km to the north and upwind of the city (Figure 1). Regarding the Shiraz sample locations, S1 is a low traffic area on the outskirts of the city and close to agricultural land to the west, S2 and S3 are residential areas of low traffic and a high density of green space, S4 is the northern entrance to the city centre and is adjacent to a major highway, and S5 is a residential area with a moderate amount of traffic.

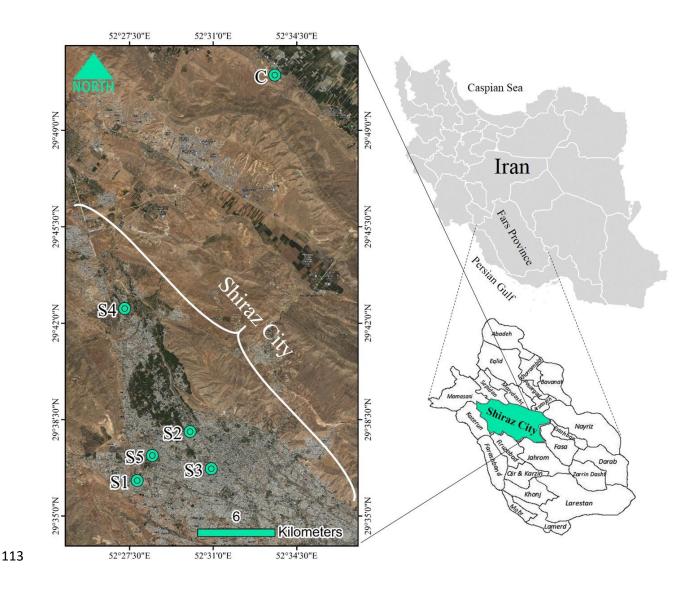


Figure 1: Location of the sampling sites in the Shiraz region. Note the relative remoteness of station C from centres of population (in grey).

# 2.2. Sampling

After carefully checking the rainfall forecasts, eight consecutive days (192 h) were selected for sampling from 28 December 2021 to 5 Jan 2022. This timescale comprised three dry events (totalling 117 h) and two distinct rainfall regimes (totalling 75 h). Dry events were sampled across 11 periods (labelled D) of twelve hours (plus any remaining time before incipient rain), and rain

events were sampled across 13 periods (labelled R) of three to eight hours depending on the duration of precipitation.

Samples were collected at each location in a stainless steel dish (whose depth was 30 cm and aperture diameter ranged from 25 to 43 cm, depending on any constraints imposed by the precise sampling location and position). Each dish was positioned, via a metallic frame, about 2 m above the flat roof of a three-storey building that was exposed in all directions, and at a total elevation of about 10 m above ground level. During dry events, dishes were partly filled with 1-L of 2 µm-filtered, distilled water to aid retention and minimise blowby of dry deposited material. At each site, material collected over the sampling period, plus filtered (< 2 μm), distilled waterrinsings of the dish, were transferred to a 2-L glass bottle with the aid of a glass funnel. During rain events, pre-cleaned (with distilled water) dry dishes were continuously used to collect wet deposited material over each period (although, strictly, bulk deposition was captured). The contents were carefully poured into a graduated glass measuring cylinder to determine rainfall volume. With a Eutech Instruments PCD650 probe, electrical conductivity (as specific conductance and a measure of rainwater chemistry) was then determined before the contents were transferred, with the dish rinsings, to a glass bottle as above.

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# 2.3. Sample processing and microplastic counting and characterisation

The contents of each glass sample bottle were vacuum filtered through a 2  $\mu$ m pore size S&S filter paper (Blue Band, grade 589/3). Filters were transferred to individual, 60 mL foil-wrapped glass beakers and digested in 35 mL of 30%  $H_2O_2$  (Arman Sina, Tehran) until bubble formation

ceased. The contents of the digestions were refiltered through clean 2  $\mu$ m filters that were subsequently air-dried at 25 °C for 48 h in a metal cabinet housed in a clean room before being stored in glass Petri dishes.

MPs on each filter were visually identified, counted and characterised (by size, shape and colour) under a stereo digital microscope (Sairan DSM3000) at up to 200 X magnification with the aid of a 250  $\mu$ m-diameter stainless steel probe and ImageJ software. Identification of MPs was based on shape, colour, size, thickness, shininess, hardness and surface structure according to protocols outlined elsewhere (Abbasi et al., 2019). Classification was based on shape (fibre, film, fragment or spherule) and length of the longest axis, L (L < 100  $\mu$ m, L > 1000  $\mu$ m, L > 1000  $\mu$ m; and with a size detection limit estimated to be between 30 and 50  $\mu$ m depending on shape). The polymeric composition of 34 MPs (or about 4% of the total) of a range of shapes, sizes and colours and collected from different locations and meteorological conditions was determined using a micro-Raman spectrometer ( $\mu$ -Raman-532-Ci, Avantes, Apeldoorn, Netherland) with a laser of 785 nm and Raman shift of 400-1800 cm<sup>-1</sup>.

# 2.4. Laboratory cleanliness and quality assurance

Laboratory equipment was washed with phosphate-free soap, double-rinsed with distilled water and soaked in 10% HNO $_3$  for 24 h before being rinsed twice with filtered, distilled water, dried at room temperature in a customised clean room and protected by Al foil. Laboratory work surfaces were cleaned with ethanol, laboratory clothing was cotton-based and all reagents and solutions were filtered through 2  $\mu$ m before being used. Under these conditions, processing of distilled water contained in glass bottles as above revealed no MP contamination. The number of MPs on

ten random (sample) filters were recounted under the microscope and returned the same values as the original counts.

# 3. Results

Table 1 shows the duration of each sampling event, along with an indication of regional wind speed, wind direction and relative humidity (obtained from ventusky.com) for each period. Average wind speeds ranged from < 5 to 10 - 15 km h<sup>-1</sup> and usually had a southerly and/or westerly component. Relative humidity exceeded 80% during the rain events but was more variable during dry events (between 10% and 90%). Also shown for each rainy period is rainfall (in mm), derived from the volume of precipitation collected and the area of the metal collecting dish, along with the measured specific electrical conductance of rainwater. For a given period, rainfall exhibited some variation among the different locations, and in particular the highest values were frequently reported for S4 to the north of the city. Overall, average rainfall intensity (i.e., mean periodic rainfall for all locations normalised to sampling duration) ranged from about 0.25 mm h<sup>-1</sup> for R6 to > 2.2 mm h<sup>-1</sup> for R18, and specific electrical conductance, as a measure of rainwater chemistry, ranged from about 20 to 270  $\mu$ s cm<sup>-1</sup>, with the highest values at each location usually encountered towards the beginning of both rain events.

Table 2 presents the number of MPs retrieved from the filters at each location and for each sampling period, along with the number normalised per m<sup>2</sup> of sampling area and per h of collection time (i.e., deposition rate). In total, 132 samples were collected and analysed (twelve were unsuccessfully retrieved or lost), and in all but eleven of these samples MPs were identified.

The number of MPs in each sample was > 20 in eleven cases, and the maximum number of MPs was 47. When normalised, deposition rates (where detected) ranged from < 1 m<sup>-2</sup> h<sup>-1</sup> (< 24 m<sup>-2</sup> d<sup>-1</sup>) in three samples taken during dry conditions to > 100 m<sup>-2</sup> h<sup>-1</sup> (> 2400 m<sup>-2</sup> d<sup>-1</sup>) in three samples collected during rain events.

Table 1: Duration of the dry (D) and rainy (R) sampling events and periods, along with indicative, average (by rank) regional wind speeds and directions and relative humidities. Periodic rainfall, derived from the volume of precipitation and area of sample capture, and specific electrical conductance (EC) of rainwater, are shown for each location. Note that NC denotes no sample collection and LS denotes a lost sample.

							С	S1	S2	S3	S4	S5
sampling period	event	duration, h	time since start, h	wind speed, km h <sup>-1</sup>	wind direction	humidity, %			rainfall, mm (EC, μS cm <sup>-1</sup> )			
D1		12	12	< 5	SW, SE	10-30						
D2	1	12	24	10-15	SW, W	10-30						
D3		3	27	< 5	S	30						
R4		3	30	< 5	SW	90	0.93 (140)	3.06 (50)	2.21 (150)	1.48 (140)	1.86 (200)	LS
R5		5	35	10-15	SW, S	90	1.36 (80)	4.39 (70)	2.65 (80)	2.55 (90)	5.00 (130)	3.49 (50)
R6		5	40	< 5	SW	90	NC	1.12 (90)	2.82 (20)	0.55 (190)	0.99 (270)	0.63 (130)
R7	2	8	48	< 5	S	90	1.12 (70)	5.10 (30)	2.92 (80)	3.55 (80)	4.65 (90)	0.93 (50)
R8		6	54	< 5	SE	90	2.44 (50)	6.53 (20)	4.91 (30)	6.14 (50)	4.94 (50)	6.40 (30)
R9		5	59	< 5	S	90	9.88 (30)	6.51 (20)	5.18 (20)	4.29 (40)	12.50 (90)	10.23 (40)
R10		5	64	5-10	SW	90	1.80 (150)	3.06 (20)	3.94 (20)	4.29 (70)	4.10 (90)	1.10 (110)
R11		8	72	< 5	W	90	NC	9.84 (20)	6.28 (20)	9.82 (20)	15.56 (170)	8.84 (50)
D12		12	84	< 5	SW	40-90		, ,	, ,	, ,	` ,	, ,
D13	3	12	96	< 5	SW	70						
D14		12	108	10-15	SW, S	40-70						
D15		6	114	< 5	SW	70						
R16		6	120	< 5	SE	80-90	1.05 (140)	3.67 (40)	1.42 (60)	2.62 (100)	2.79 (170)	7.05 (30)
R17		6	126	10-15	SE, S	90-100	8.72 (20)	12.76 (20)	7.08 (60)	12.44 (40)	10.16 (70)	9.53 (30)
R18	4	5	131	< 5	Š	100	NC	ĹS	9.69 (20)	7.24 (40)	20.70 (40)	6.40 (50)
R19		5	136	< 5	S	90-100	5.12 (20)	LS	2.43 (30)	3.21 (80)	7.91 (70)	2.73 (30)
R20		8	144	< 5	S, W	80-90	7.38 (50)	LS	6.33 (30)	5.06 (70)	11.63 (70)	7.56 (50)
D21		12	156	5-10	SW	50-70	. ,		,	. ,	, ,	,
D22	5	12	168	5-10	SW, S	80						
D23	-	12	180	10-15	NW, W	40-80						
D24		12	192	< 5	SW	60						

Table 2: The number of MPs and the number of MPs per m<sup>2</sup> per h for each sampling event and sampling period and for each location. Note that NC denotes no sample collection and LS denotes a lost sample.

		no. MPs								MP m <sup>-2</sup> h <sup>-1</sup>				
sampling period	event	С	S1	S2	S3	S4	S5	_	С	S1	S2	<b>S</b> 3	S4	S5
D1		6	4	13	9	12	4		5.8	6.8	9.6	5.2	11.6	3.9
D2	1	9	11	18	21	20	15		8.7	18.7	13.3	12.1	19.4	14.5
D3		6	3	15	7	12	3		23.3	20.4	44.2	16.1	46.5	11.6
R4		4	19	20	12	26	LS		15.5	129.3	59.0	27.6	100.8	LS
R5		44	10	11	12	25	20		102.3	40.8	19.5	16.6	58.1	46.5
R6		NC	6	3	2	1	12		NC	24.5	5.3	2.8	2.3	27.9
R7	2	13	5	1	4	8	2		18.9	12.8	1.1	3.4	11.6	2.9
R8		17	6	2	5	2	3		32.9	20.4	2.9	5.7	3.9	5.8
R9		4	4	2	4	0	1		9.3	16.3	3.5	5.5	0	2.3
R10		2	6	2	10	18	3		4.7	24.5	3.5	13.8	41.9	7.0
R11		NC	1	1	2	14	3		NC	2.6	1.1	1.7	20.3	4.4
D12		0	0	1	1	4	1		0	0	0.7	0.6	3.9	1.0
D13	3	1	0	2	5	LS	6		1.0	0	1.5	2.9	LS	5.8
D14		0	1	4	8	4	7		0	1.7	2.9	4.6	3.9	6.8
D15		4	2	0	0	2	10		7.8	6.8	0	0	3.9	19.4
R16		5	5	8	5	13	14		9.7	17.0	11.8	5.7	25.2	27.1
R17		14	3	9	20	18	1		27.1	10.2	13.3	23.0	34.9	1.9
R18	4	NC	LS	2	47	3	3		NC	LS	3.5	64.8	7.0	7.0
R19		1	LS	8	2	2	1		2.3	LS	14.2	2.8	4.7	2.3
R20		2	LS	11	5	1	1		2.9	LS	12.2	4.3	1.5	1.5
D21		1	1	4	1	1	0		1.0	1.7	2.9	0.6	1.0	0
D22	5	4	LS	0	8	4	0		3.9	LS	0	4.6	3.9	0
D23		2	LS	9	20	LS	7		1.9	LS	6.6	11.5	LS	6.8
D24		5	LS	10	21	5	0		4.8	LS	7.4	12.1	4.8	0

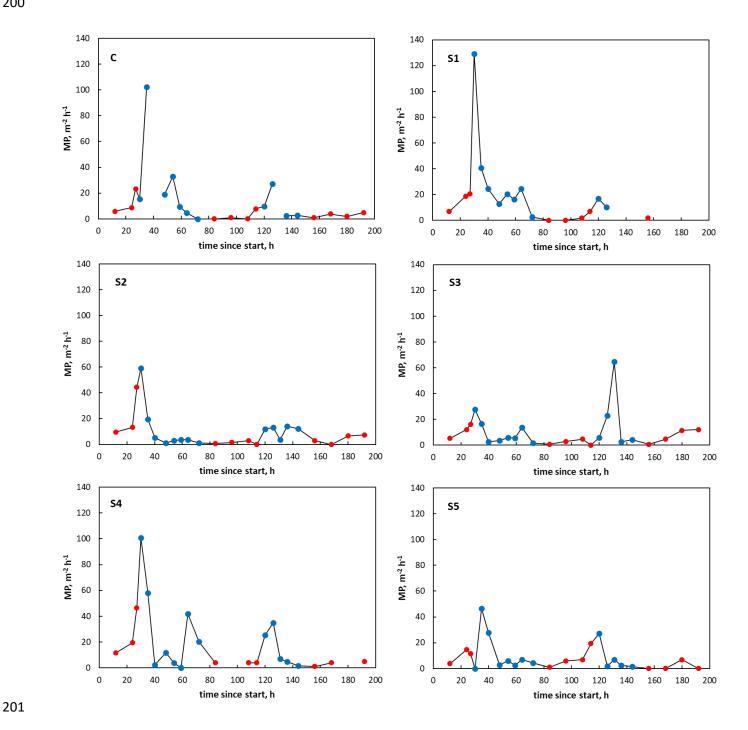


Figure 2: MP deposition as a function of time at the six locations (C and S1 to S5). Dry events are shown in red and rain events are shown in blue. Note the discontinuities where samples were lost or not collected (see caption to Figure 2).

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The evolution of MP deposition over the entire timeframe studied is illustrated for all six locations in Figure 2. Distributions were broadly similar at each location and consisted of an initial dry event in which deposition increased, followed by a rain event in which deposition rapidly peaked and subsequently declined but with evidence of a secondary peak whose precise timing was more variable. Low MP deposition was observed in the subsequent dry event, which exhibited an increase at three locations, before the second precipitation event in which deposition exhibited a persistent peak of variable magnitude and timing. During the final dry event, deposition was again low, and in most cases exhibited an increase as a function of time. With the exception of seven fragments (sampled in both dry and rainy conditions), all MPs detected in the study were fibres of varying lengths, thicknesses and colours. The size distributions of MPs, combined for all locations, are shown in Figure 3 by sampling event (defined in Table 1). For all dry and rain events, the largest (> 1000 μm) and smallest (< 100 μm) size categories were most important, with a combined proportion that averages about twothirds, and remaining MPs were distributed among the three intermediate size classifications. The results of the Raman analysis are summarised in Figure 4. Here, 31 fibres and three fragments were analysed from various locations and events. Polyethylene terephthalatepolyester was the dominant polymer (> 50%), with additional contributions from polyvinyl chloride, polytetrafluoroethylene, polystyrene, polypropylene, polyethylene and nylon.

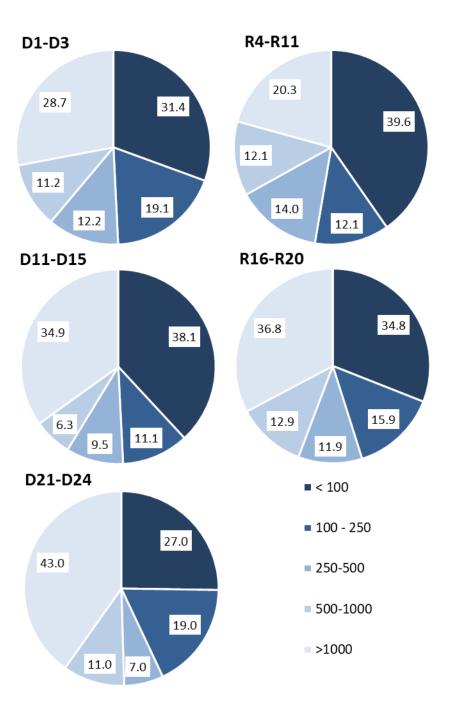


Figure 3: Percentage distribution of MPs by size (L, in  $\mu$ m) for five different dry (D) and rain (R) events (see Table 1).

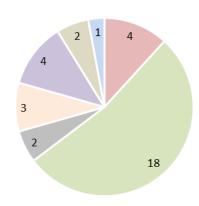




Figure 4: Distribution of MPs by polymer type (n = 34). PET = polyethylene terephthalate, PTFE = polytetrafluoroethylene, and numbers in parentheses in the legend denote non-fibrous samples (fragments).

# 4. Discussion

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Based on our sampling design and identification protocols, the broad, qualitative findings of the 266 present study are in agreement with those from other studies examining the presence or 267 deposition of atmospheric MPs in urban and more remote areas, and either in dry conditions or during rainfall or both. Here, fibres of varying lengths are the most abundant shape of MP, presumably because of their favourable aerodynamic properties, and PET-polyester is the most 270 common (or one of most common) petroleum-based polymer/s (Wright et al., 2020; Abbasi and 271 Turner, 2021; Ding et al., 2021; Huang et al., 2021; Purwiyanto et al., 2021; Jia et al., 2022; Liu et al., 2022; Perera et al., 2022; Chang et al., 2023; Li et al., 2023). The dominance of polyester-272 273 based fibres suggest that clothing and textiles are the most important primary source of 274 airborne MPs, with secondary distal sources likely to include resuspension from soils and agricultural land applied with contaminated sewage sludge (Brahney et al., 2020; Rezaei et al., 275 276 2022). Atmospheric deposition rates of MPs reported in the literature vary widely. For example, 277 Purwiyanto et al. (2022) determined fluxes between 3 and 40 MPs m<sup>-2</sup> d<sup>-1</sup> in coastal Jakarta. while Jia et al. (2022) report an average of 3261 MPs m<sup>-2</sup> d<sup>-1</sup> in Shanghai during the rainy season and Li et al. (2023) report a maximum of 75,000 MPs m<sup>-2</sup> d<sup>-1</sup> in a rural region of northern China. Geographical and temporal variations have been attributed to seasonal and climatic conditions 282 (Huang et al., 2021; Hernández-Fernández et al., 2022) and proximity to urban areas (Shruti et al., 2022; Sun et al., 2022), as well as differences in sample collection (including height above 283 ground level) and processing and means of plastic identification (Knobloch et al., 2021; Szewc et al., 2021). For example, we note that Li et al. (2023) sampled at only 100 cm above ground level 285 and report rayon as the dominant polymer type, indicating that semi-synthetic fibres were considered alongside petroleum-based polymers. 288 One of the key factors determining the nature and magnitude of atmospheric MP depositional 289 flux is the mode of deposition (dry deposition versus wet deposition). Dry deposition is determined by deposition velocity (particle size and shape) and is also specific to meteorological conditions and landcover type (Klein et al., 2019; Szewc et al., 2021), while 291

rainfall acts to scavenge MPs from the atmosphere (Zhang et al., 2020). Many studies have found that, for a given sampling time, wet deposition is greater than dry deposition, or at least bulk deposition is greater during wet periods than dry periods (Dris et al., 2016; Huang et al., 2021; Liu et al., 2022; Li et al., 2023; Yuan et al., 2023). However, these studies have only considered daily or cumulative rainfall and have acknowledged that other factors, like rainfall height, intensity and frequency and droplet size, could be important. The potential effects of rainfall intensity and duration on MP deposition were demonstrated by Abbasi (2021) during the onset of a monsoon event in Shiraz. Here, MP deposition rate rapidly declined during the first 30 minutes of heavy precipitation. The present study has determined MP deposition with a relatively high temporal resolution (several hours, and with MPs reported per m<sup>-2</sup> per h) over consecutive dry and rain events and, therefore, affords a greater insight into the role that rainfall plays in MP transport. Thus, across six locations within the vicinity of Shiraz that includes a more remote location, and despite different land uses and local sources of MPs, distributions and magnitudes of deposition across the entire sampling timeframe are similar. Specifically, an initial dry weather event, with southerly winds and low humidity, allows MPs from local and distal sources to progressively build up in the atmosphere and deposit through gravitational settlement. Incipient rainfall acts to rain out and wash out these MPs (through incloud and below-cloud scavenging, respectively; Audoux et al., 2023), resulting in a spike in wet deposition with little or no lag. Subsequent precipitation encounters "cleaner" air and depositional fluxes decline to a low, residual value. A secondary peak in deposition could be related to a spell of relatively high rainfall (and, presumably, relatively large raindrops that increase collision frequencies; Guo et al., 2016) coupled with winds with an easterly component (Table 1) that introduce MPs from a different region. Specific electrical conductance follows these broad trends, reflecting the propensity of rainfall to concurrently washout airborne salts from the atmosphere. The ensuing dry event is accompanied by an immediate period of low deposition (initially, no MPs were detected at C and S1), and although there is often a subsequent build-up of MPs, the

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increase is not as marked as that observed during the initial dry event. Presumably, this reflects the effects of preceding rainfall in cleansing atmospheric MPs and the wetting of regional soils that inhibits MP resuspension (Abbasi and Turner, 2021). An additional contributing factor could be that rainfall here began in the early hours of the next day when local MP-generating activities were minimal. The next rainfall event is accompanied by a smaller peak that reflects the washing out of lower numbers of locally resuspended and regionally advected MPs, with the final dry event beginning with very little MP deposition and exhibiting evidence that the preceding cycle is repeated.

The intra- and inter-event patterns that we observed for MPs, and in particular those during incipient rainfall, have been documented for other aerosols more generally (Battarbee et al., 1997; Castro et al., 2010; Fujino and Miyamoto, 2022). Here, washout is usually reported to be greater for larger particles, and although this effect was not evident from our results, it must be borne in mind that the more general literature deals with much smaller and higher density particulate matter. Consistent with other independent studies, rainfall acts to remove MPs from the atmosphere, at least temporarily.

In summary, average MP depositional fluxes were similar among multiple locations with different characteristics within the Shiraz region, and significant (at least order of magnitude) intra-location variations observed during different meteorological conditions. These findings support previous assertions that MPs are ubiquitous contaminants, but also suggest that variations reported in the literature may result, to a significant extent, from environmental conditions and not, necessarily, proximity to source regions. Specifically, it would appear that incipient rainfall is a critical driver for the scavenging of MPs from the atmosphere through below-cloud washout. It is surmised that rainfall also acts to inhibit MP resuspension from certain surfaces (e.g., soils) through ground dampening.

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#### 351 **References**

- Abbasi, S., 2021. Microplastics washout from the atmosphere during a monsoon rain event.
- Journal of Hazardous Materials Advances 4, 100035.
- Abbasi, S., Turner, A., 2021. Dry and wet deposition of microplastics in a semi-arid region
- 355 (Shiraz, Iran). Science of the Total Environment 786, 147358.
- Allen, S., Allen, D., Phoenix, V.R., Roux, G.L., Jim'enez, P.D., Simonneau, A., Binet, S., Galop, D.,
- 357 2019. Atmospheric transport and deposition of microplastics in a remote mountain catchment.
- 358 Nature Geosciences 12, 339–344.
- Audoux, T., Laurent, B., Chevallier, S., Feron, A., Pangui, E., Maisonneuve, F., Desboeufs, K.,
- 360 Triquet, S., Noyalet, G., Lauret, O., Huet, F., 2023. Automatic sequential rain sampling to study
- atmospheric particulate and dissolved wet deposition. Atmospheric Environment 295, 119561.
- Battarbee, J.L., Rosel, N.L., Long, X., 1997. A continuous, high resolution record of urban
- 363 airborne particulates suitable for retrospective microscopical analysis. Atmospheric
- 364 Environment 31, 171-181.
- Brahney, J., Hallerud, M., Heim, E., Hahnenberger, M., Sukumaran, S., 2020. Plastic rain in
- protected areas of the United States. Science 368, 1257–1260.
- Castro, A., Alonso-Blanco, E., González-Colino, M., Calvo, A.I., Fernández-Raga, M., Fraile, R.,
- 368 2010. Aerosol size distribution in precipitation events in León, Spain Atmospheric Research 96,
- 369 421-435.
- 370 Chang, D.Y., Jeong, S., Shin, J., Park, J., Park, C.R., Choi, S., Chun, C.H., Chae, M.Y., Lim, B.C.,
- 371 2023. First quantification and chemical characterization of atmospheric microplastics observed
- in Seoul, South Korea. Environmental Pollution 327, 121481.
- Ding, Y., Zou, X., Wang, C., Feng, Z., Wang, Y., Fan, Q., Chen, H., 2021. The abundance and
- 374 characteristics of atmospheric microplastic deposition in the northwestern South China Sea in
- the fall. Atmospheric Environment 253, 118389.
- Dris, R., Gasperi, J., Saad, M., Mirande, C., Tassin, B., 2016. Synthetic fibers in atmospheric
- fallout: a source of microplastics in the environment? Marine Pollution Bulletin 104, 290–293.
- Fujino, R., Miyamoto, Y., 2022. PM<sub>2.5</sub> decrease with precipitation as revealed by single-point
- 379 ground-based observation. Atmospheric Science Letters 23, e1088.
- 380 Guo, L.C., Zhang, Y., Lin, H., Zeng, W., Liu, T., Xiao, J., Rutherford, S., You, J., Ma, W., 2016. The
- 381 washout effects of rainfall on atmospheric particulate pollution in two Chinese cities.
- 382 Environmental Pollution 215, 195-202.

- 383 Hernández-Fernández, J., Puello-Polo, E., Trillera, J., 2022. Characterization of microplastics in
- total atmospheric deposition sampling from areas surrounding industrial complexes in
- northwestern Colombia. Sustainability 14, 13613.
- 386 Huang, Y., He, T., Yan, M., Yang, L., Gong, H., Wang, W., Qing, X., Wang, J., 2021. Atmospheric
- transport and deposition of microplastics in a subtropical urban environment. Journal of
- 388 Hazardous Materials 416, 126168.
- 389 Jia, Q., Duan, Y., Han, X., Sun, X., Munyaneza, J., Ma, J., Xiu, G., 2022. Atmospheric deposition of
- microplastics in the megalopolis (Shanghai) during rainy season: Characteristics, influence
- factors, and source. Science of the Total Environment 847, 157609.
- Klein, M., Fischer, E.K., 2019. Microplastic abundance in atmospheric deposition within the
- Metropolitan area of Hamburg, Germany. Science of the Total Environment 685, 96–103.
- Knobloch, E., Ruffell, H., Aves, A., Pantos, O., Gaw, S., Revell, L.E., 2021. Comparison of
- deposition sampling methods to collect airborne microplastics in Christchurch, New Zealand.
- 396 Water Air and Soil Pollution 232, 133.
- Li, J., Zhang, J., Ren, S., Huang, D., Liu, F., Li, Z., Zhang, H., Zhao, M., Cao, Y., Mofolo, S., Liang, J.,
- 398 Xu, W., Jones, D.L., Chadwick, D.R., Liu, X., Wang, K., 2023. Atmospheric deposition of
- microplastics in a rural region of North China Plain. Science of the Total Environment 877,
- 400 162947.
- 401 Liu, Z., Bai, Y., Ma, T., Liu, X., Wei, H., Meng, H., Fu, Y., Ma, Z., Zhang, L., Zhao, J., 2022.
- 402 Distribution and possible sources of atmospheric microplastic deposition in a valley basin city
- 403 (Lanzhou, China). Ecotoxicology and Environmental Safety 233, 113353.
- 404 Perera, K., Ziajahromi, S., Nash, S.B., Manage, P.M., Leusch, F.D.L., 2022. Airborne microplastics
- in indoor and outdoor environments of a developing country in South Asia: Abundance,
- 406 distribution, morphology, and possible sources. Environmental Science & Technology 56,
- 407 16676-16685.
- 408 Purwiyanto, A.I.S., Prartono, T., Riani, E., Naulita, Y., Cordova, M.R., Koropitan, A.F., 2022. The
- deposition of atmospheric microplastics in Jakarta-Indonesia: The coastal urban area. Marine
- 410 Pollution Bulletin 174, 113195.
- 411 Rezaei, M., Abbasi, S., Pourmahmood, H., Oleszczuk, P., Ritsema, C., Turner, A., 2022.
- 412 Microplastics in agricultural soils from a semi-arid region and their transport by wind erosion.
- 413 Environmental Research 212, 113213.
- 414 Shruti, V.C., Kutralam-Muniasamy, G., Peréz-Guevara, F., Roy, P.D., Martínez, I.E., 2022.
- 415 Occurrence and characteristics of atmospheric microplastics in Mexico City. Science of the Total
- 416 Environment 847, 157601.

- Sun, J., Peng, Z., Zhu, Z.R., Fu, W., Dai, X., Ni, B.J., 2022. The atmospheric microplastics
- deposition contributes to microplastic pollution in urban waters. Water Research 225, 119116.
- 419 Szewc, K., Graca, B., Dołega, A., 2021. Atmospheric deposition of microplastics in the coastal
- 420 zone: Characteristics and relationship with meteorological factors. Science of the Total
- 421 Environment 761, 143272.

433

- Villanova-Solano, C., Hernández-Sánchez, C., Díaz-Peña, F.J., González-Sálamo, J., González-
- 423 Pleiter, M., Hernández-Borges, J., 2023. Microplastics in snow of a high mountain national park:
- 424 El Teide, Tenerife (Canary Islands, Spain). Science of the Total Environment 873, 162276.
- Wright, S.L., Ulke, J., Font, A., Chan, K.L.A., Kelly, F.J., 2020. Atmospheric microplastic deposition
- in an urban environment and an evaluation of transport. Environ. Int. 136, 105411.
- 427 Yuan, Z., Pei, C.L., Li, H.X., Lin, L., Liu, S., Hou, R., Liao, R., Xu, X.R., 2023. Atmospheric
- 428 microplastics at a southern China metropolis: Occurrence, deposition flux, exposure risk and
- washout effect of rainfall. Science of the Total Environment 869, 161839.
- 430 Zhang, Y., Kang, S., Allen, S., Allen, D., Gao, T., Sillanpää, M., 2020. Atmospheric microplastics: a
- review on current status and perspectives. Earth Science Reviews 203, 103118.